

First record of a fish-killing *Gymnodinium* sp. bloom in Kuwait Bay, Arabian Sea: chronology and potential causes

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ABSTRACT: Significant natural and aquaculture fish deaths in Kuwait Bay occurred from September to October of 1999 and were attributed to a bloom of the dinoflagellate *Gymnodinium* sp. A chronology of the bloom event suggests that a period of low winds and stable water-column structure preceded the bloom. Maximum cell concentrations of *Gymnodinium* sp. ($>6 \times 10^6$ cells l⁻¹) were also immediately preceded by a more than 20-fold increase in mean inorganic nitrogen concentrations (up to 60 μM) and elevated inorganic phosphate concentrations. This, combined with elevated inorganic and organic nutrient concentrations within the bloom, suggests that coastal nutrient eutrophication was likely to have contributed significantly to bloom development and support. Termination of the *Gymnodinium* sp. bloom coincided with a bloom of the non-toxic ciliate *Mesodinium rubrum*, which appeared as large red patches in Kuwait Bay. While no adverse human health effects were associated with the bloom, closure of shellfish and selected finfish (largely mullet *Liza macrolepis*) markets resulted in significant economic losses to the region. The occurrence of this toxic algal bloom event, the first within the Arabian Sea, highlights the need for monitoring and research programs in the Arabian Sea and Kuwait Bay that focus on nutrients and eutrophication, in addition to oil related pollution issues.

KEY WORDS: Arabian Sea · Kuwait Bay · *Gymnodinium* sp. · *Mesodinium rubrum* · Eutrophication · Fish kill · Nitrogen · Nutrients · Red tide

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INTRODUCTION

The Arabian Gulf is a shallow, semi-enclosed marginal sea characterized by both extreme natural environmental conditions and severe anthropogenic stresses. Evaporation exceeds combined rainfall and coastal freshwater inputs to this region by a factor of 10 (Sheppard 1993). This, combined with an average flushing time of 3 to 5 yr (Hunter 1986), results in salinities of up to 44.30‰ (Saad 1976, Jacob & Al-Muzaini

1990). Water temperatures vary from 12 to >32°C (ROPME 1999). Superimposed on these extreme environmental conditions are a series of unique environmental pressures. Oil-related pollution inputs to the Gulf, especially the northern Gulf, are immense. Pre-Gulf War inputs alone are estimated to have been 47 times that of the remaining total estimated global oil pollution (Golob & Bruss 1984), and estimates of direct Gulf War-related oil inputs range from 6 to 12 million barrels (Readman et al. 1992, Cava et al. 1993, Litherathy 1993). In the northern Gulf, major point-source pollutant inputs include sewage outfalls from Kuwait City and outfalls associated with industrial, desalina-

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tion and power plant complexes to the north of Kuwait City. Additionally, a large drainage channel (anecdotally known as the 'third river'), constructed after the 1991 Gulf War by Iraq, drains the southwest Al-Ahwar wetlands of coastal Iraq at the Iraq-Kuwait border. This channel has resulted in a localized reduction in salinity and has shifted the source of nutrient inputs from Shatt Al-Arab directly into Kuwaiti waters. The long-term consequences of this change are unknown (ROPME 1999).

Subba Rao & Al-Yamani (1998) have summarized available knowledge of phytoplankton from the northern Arabian Gulf and described the region as a distinct biotope, with a north-south gradient in phytoplankton diversity and abundance. Populations within the Tigris and Euphrates River estuaries are characterized by greater biomass, but are less diverse than those of open Gulf waters to the south. Little information is available on phytoplankton rate processes within the region beyond infrequent primary production measurements (Al-Abaychi & Ghani 1986, Hadi et al. 1989, Sheppard et al. 1992, Subba Rao & Al-Yamani 1998).

The occurrence of harmful algal blooms is often cited as symptomatic of increasing coastal environmental stresses, including eutrophication, pollution and increases in aquaculture (see review in Hallegraeff 1995). Available phytoplankton species lists from the Arabian Sea (see Subba Rao & Al-Yamani 1998 for summary) suggest that some potentially harmful species are present within the region. In September of 1999, a major fish kill occurred in Iranian coastal waters and was immediately followed by wild and aquaculture-related fish kills in Kuwaiti coastal waters. Mortality (at least in Kuwait waters) was related to the presence of a red tide outbreak of the dinoflagellate *Gymnodinium* sp., previously identified only from New Zealand coastal waters (MacKenzie et al. 1995, 1996, Haywood, Steidinger, Truby & Mackenzie unpubl.). This as yet undescribed species is being named '*selliformis*' after its saddle-shaped hypothecal notch (Haywood et al. unpubl.) We report here a chronology of the bloom, environmental conditions during the bloom, potential causes, and the implications for environmental monitoring in the Arabian Gulf.

METHODS

Kuwait Bay is a large (~850 km²), sub-tropical, semi-enclosed embayment located in the northwest region of the Arabian Sea (Fig. 1). The Bay is shallow, with a maximum depth of 20 m and a mean depth of <3 m (Anderlini et al. 1982), so most of it lies within the photic zone year-round (Sheppard 1993, Carpenter et al. 1997). The tidal range is >1 m, and mixing and

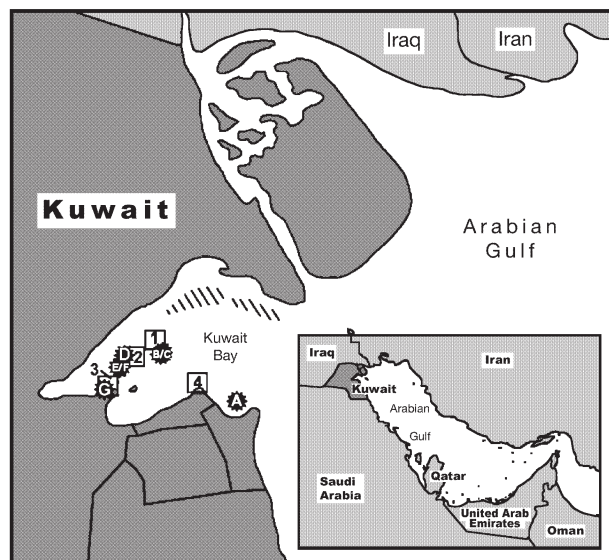


Fig. 1. Map of Kuwait Bay. Stns 1 to 4 were sampled on October 14, 1999 and represent the Kuwait Environment Public Authority (EPA) long-term monitoring stations. Stns A to G were sampled on October 15, 1999. Hatched area is the region which experienced fish kill of mullet on September 19, 1999

circulation are largely wind-driven (Lehr 1984, Abou-Seida & Al-Sarawi 1990, Reynolds 1993).

Samples for phytoplankton community composition and nutrient analyses were collected during the Kuwait red tide outbreak on October 14 and 15, 1999, at a range of sites located throughout Kuwait Bay (Fig. 1). Four stations sampled from 10:00 to 12:00 h on October 14 coincided with long-term monitoring stations maintained by the Kuwait Environment Public Authority (EPA). Seven additional stations, selected to cover a range of environmental conditions including fish-kill areas, aquaculture cages and areas adjacent to major industrial outfalls, were sampled on October 15 from 09:00 to 12:00 h.

Water samples were collected with a modified Niskin bottle from 10 cm below the water surface, stored in a shady area on the boat, and returned to the laboratory within 3 h of collection. Nutrient samples were immediately filtered onto precombusted GF/F filters (450°C for 1 h), frozen, and later transported on ice to Horn Point Laboratory, University of Maryland, Cambridge, USA, for analysis. The filters were retained for chlorophyll a analysis (Parsons et al. 1984). All samples remained frozen during transport. Inorganic nutrients [NO_3^- , NH_4^+ , PO_4^{3-} and Si(OH)_4] were analyzed using standard autoanalyzer methods (Parsons et al. 1984). Samples for total dissolved nitrogen were analyzed on Antek Instruments high-temperature oxidation instruments, calibrated using a urea standard (Bronk et al. 2000). Concentrations of dissolved organic nitrogen

(DON) were calculated by subtracting all inorganic nitrogen concentrations from the total dissolved nitrogen concentrations. Samples for total dissolved phosphate were oxidized using the persulfate oxidation method (Valderrama 1981) and analyzed on an autoanalyzer as described above. Dissolved organic phosphorus (DOP) concentrations were calculated as the difference between total dissolved phosphate and PO_4^{3-} . Samples for dissolved organic carbon (DOC) were analyzed on a Shimadzu TOC instrument. Additional unpublished inorganic nutrient data from the months immediately prior to the bloom were made available by the Kuwait monitoring program.

Gymnodinium sp. and *Mesodinium rubrum* concentrations were determined prior to sample preservation within 1 h of return to the laboratory. Samples were mixed by gently inverting the sample container twice, then duplicate 0.1 ml volumes from each sample were counted using a Palmer-Maloney counting chamber and an Olympus BH-2 microscope according to Thronsdon (1995). Approximately 100 ml of water from each station was preserved with Lugol's solution for subsequent analysis of phytoplankton composition with an inverted microscope according to Hasle (1978). Community composition was determined to species level when possible, and size measurements were made using an Olympus ocular micrometer.

RESULTS

A chronology of the bloom event is given in Table 1. A fish kill in Iranian waters preceded by more than 1 mo the first reported fish kill (primarily mullet, *Liza macrolepis*), which was observed in Kuwaiti waters on September 19. The direct cause of the fish kill off Iran was not documented. Discolored water was observed in both northern and southern Kuwait Bay throughout the week following this initial large fish kill, and a large fish kill (sobaity, *Acanthopagrus cuvieri*, 80 000 fish) in aquaculture enclosures in northern Kuwait Bay occurred on October 1 and 2. A bloom of a gymnodinoid dinoflagellate (Fig. 2) was identified as the cause of the fish kills by Kuwaiti EPA personnel on October 2 based upon high ($>6 \times 10^6$ cells l^{-1}) *Gymnodinium* sp. cell concentrations in areas with fish kills and the elimination of other potential causes (e.g. accidental discharge of chemical pollutants, oil and trace metals) via intensive measurements made subsequent to the first fish kill (Kuwait EPA 1999). Following the peak of the *Gymnodinium* sp. bloom, red water caused by *Mesodinium rubrum* was observed (Table 2).

Some confusion over the identification of the *Gymnodinium* existed during the initial stages of the bloom, as this species exhibited some of the morphological

characteristics of both *Gymnodinium mikimotoi* (i.e. straight apical groove, cell size and girdle displacement) and *G. breve* (i.e. pronounced hypotheca invagination); both 'large' and 'small' vegetative cells (see Silva & Faust 1995) were observed in bloom populations (cell width 13 to 35 μm , cell length 17 to 37 μm). Although initially identified as *G. cf. mikimotoi* using light microscopy, the *Gymnodinium* species responsible for the Kuwait red tide was confirmed to be the type found in New Zealand that is being described as '*selliformis*' which is similar to *G. mikimotoi* but morphologically and genetically distinct (Haywood et al. unpubl.). Identification was subsequently confirmed as this new *Gymnodinium* species by A. Haywood

Table 1. Chronology of *Gymnodinium selliforme* bloom and associated toxicity events in Kuwait in 1999

Date	Observations
15 Sep	Fish kill (~15 t, 1500 km ²) along coast of Khuzestan Province, Iran
19 Sep	Fish kill (~25 to 30 t), primarily mullet <i>Liza macrolepis</i> in northern Kuwait Bay reported by Kuwait EPA
20 Sep	Fish kill (<i>L. macrolepis</i>) in nearshore, northern and southern Kuwait Bay. Kuwait EPA informed Regional Organization for the Protection of the Marine Environment (ROPME) of the event
21 Sep	Patches of discolored water observed south of Kuwait Bay
28–29 Sep	Discolored water observed off El-Bedee (Salmiya) Beach by overflight
1–2 Oct	Fish kill (~80 000 fish) of sobaity seabream <i>Acanthopagrus cuvieri</i> in mariculture farms near Doha, Kuwait Bay
3 Oct	Discolored water observed in Kuwait Bay, continuation of fish death in mariculture farms
4 Oct	Dinoflagellate species causing discolored water identified as <i>Gymnodinium</i> sp. by Kuwait EPA, with cell concentrations $>6 \times 10^6$ cells l^{-1}
12 Oct	Patches of discolored water observed in northern and southern Kuwait Bay with small (<100 fish) daily fish kills in aquaculture enclosures; <i>Gymnodinium</i> sp. dominated phytoplankton community in Kuwait Bay (cell concentrations $<0.5 \times 10^6$ cells l^{-1}), with <i>Mesodinium rubrum</i> present
14 Oct	Patches of discolored water observed. Samples contained a mixture of <i>Gymnodinium</i> sp. and <i>M. rubrum</i> , with <i>M. rubrum</i> dominating in discolored patches
15 Oct	Patches of discolored water observed. <i>Gymnodinium</i> sp. dominated at aquaculture sites with fresh fish kill (<1000 fish). At other sites, <i>M. rubrum</i> dominated phytoplankton, with <i>Gymnodinium</i> sp. present. Red water due to <i>M. rubrum</i>
17 Oct	Kuwait EPA advises against consumption of both shellfish and fish guts and heads for a minimum of 30 d

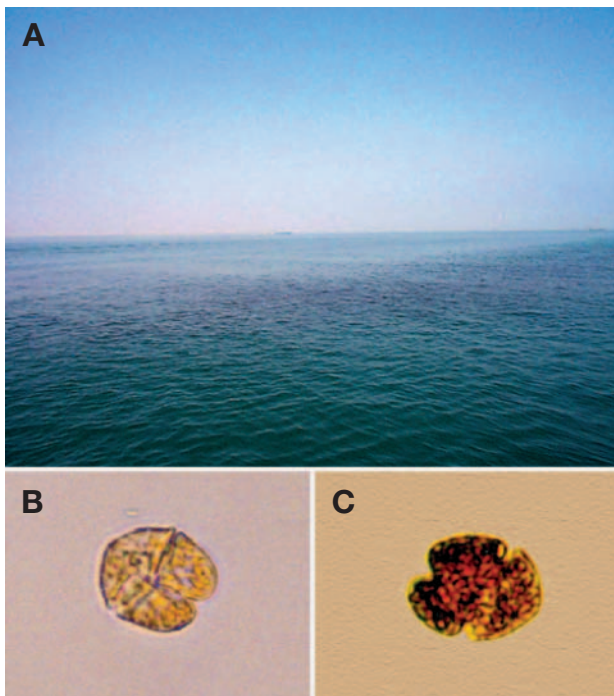


Fig. 2. (A) Red water in Kuwait Bay, October 14, 1999, (B) live *Gymnodinium selliforme* cells and (C) *G. selliforme* cells preserved with Lugol's solution

(Cawthron Institute, Nelson, New Zealand) and K. Steidinger (Florida Fish and Wildlife Conservation Commission, Florida Marine Research Institute, St. Petersburg, Florida) in December 1999, based on sample material provided by the Kuwait EPA. No *G. breve* was identified in any of the samples examined. Although other potentially toxic phytoplankton species were present within the area during the latter stages of the bloom (Table 2), none except *Gymnodinium* sp. were present at cell concentrations sufficient to cause harmful effects.

Limited aquaculture-related fish kills (<100 fish d⁻¹ cage⁻¹) continued through the week of October 16. On October 17, the Kuwait EPA issued an alert advising against the consumption of shellfish and fish guts and heads for at least 30 d. This advisory was finally lifted in March 2000. Local consumption of shellfish and finfish was greatly reduced during the initial stages of the bloom, as toxic algal

blooms were previously unknown in this region. No adverse human effects were associated with this bloom. The closure of shellfish and selected finfish markets during the bloom resulted in significant economic losses to local markets (Elizabeth 1999).

Gymnodinium sp. concentrations in samples from October 4 exceeded 6×10^6 cells l⁻¹ in the area of fish kills. Analysis of phytoplankton community composition on October 14 at 4 established Kuwait EPA monitoring stations showed that *Gymnodinium* sp. dominated phytoplankton abundance at all stations except Stn 1, where high concentrations of the diatom *Nitzschia* sp. were present (Fig. 3A). A more comprehensive survey conducted the following day, which included principal outfall and aquaculture sites within Kuwait Bay, found high (5.0×10^5 cells l⁻¹) concentrations of *Gymnodinium* sp. at aquaculture pens associated with fresh fish kills. The ciliate *Mesodinium rubrum* was present at high concentrations during this survey (up to 5.7×10^5 cells l⁻¹) at the primary industrial outfall site and within localized patches of reddish discolored water offshore of the beaches in southern Kuwait Bay (Fig. 3B).

No deviations from the normal seasonal range of physical environmental conditions was apparent in the Kuwait EPA monitoring data of Kuwait Bay prior to or

Table 2. List of phytoplankton species present in October 14 and 15 samples from Kuwait Bay during bloom event

Diatoms	Dinoflagellates	Others
<i>Asterionellopsis glacialis</i>	<i>Amphidinium carterae</i>	<i>Phaeocystis</i> sp.
<i>Bacteriastrum furcatum</i>	<i>Ceratium symmetricum</i>	<i>Mesodinium rubrum</i>
<i>Ceratiulina pelagica</i>	<i>Ceratium fusus</i>	
<i>Chaetoceros costatus</i>	<i>Dinophysis caudata</i>	
<i>Chaetoceros curvisetus</i>	<i>Dinophysis miles</i>	
<i>Chaetoceros debilis</i>	<i>Gymnodinium galantum</i>	
<i>Chaetoceros eibonii</i>	<i>Gymnodinium simplex</i>	
<i>Chaetoceros lorentzianus</i>	<i>Gymnodinium</i> sp. ^a	
<i>Chaetoceros 'b' sp.</i>	<i>Gyrodinium fusiformis</i>	
<i>Chaetoceros</i> spp.	<i>Gyrodinium spirale</i>	
<i>Coscinodiscus</i> sp.	<i>Gyrodinium</i> spp.	
<i>Cylindrotheca clostridium</i>	<i>Polykrikos</i> sp.	
<i>Eucampia zoodiacus</i>	<i>Prorocentrum concavum</i>	
<i>Guidinardia flaccida</i>	<i>Prorocentrum emarginatum</i>	
<i>Gyrosigma</i> sp.	<i>Prorocentrum lima</i>	
<i>Helicotheca tamesis</i>	<i>Prorocentrum micans</i>	
<i>Leptocylindrus</i> sp.	<i>Prorocentrum minimum</i>	
<i>Navicula</i> spp.	<i>Protoperidinium minutum</i>	
<i>Nitzschia</i> spp.	<i>Protoperidinium</i> spp.	
<i>Paralia sulcata</i>	<i>Pyrophacus steinii</i>	
<i>Pleurosigma</i> sp.	<i>Scripsiella troichiodea</i>	
<i>Pseudonitzschia</i> sp.		
<i>Rhizosolenia chunii</i>		
<i>Rhizosolenia setigera</i>		
<i>Skeletonema costatum</i>		
<i>Thalassionema nitzschoides</i>		
<i>Thalassiosira pseudonana</i>		

^aPresent as both 'large' and 'small' vegetative cells

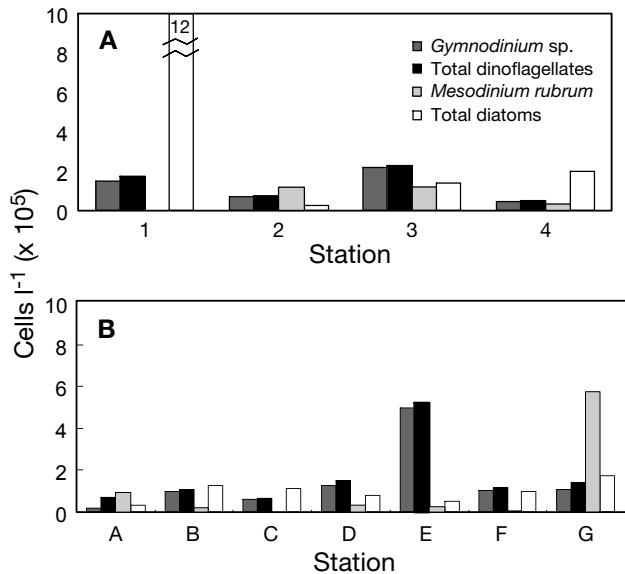


Fig. 3. Composition of the phytoplankton community for stations sampled on (A) October 14, 1999 and (B) October 15, 1999

during the bloom. Temperature ranged from 26.9 to 28.6°C, while salinity was consistently elevated at all stations sampled (from 41.32 to 42.59‰). Dissolved oxygen concentrations ranged from 3.07 to 5.05 mg l⁻¹ and displayed no relationship with dinoflagellate concentration. Secchi depths were reduced compared with pre-bloom data (Kuwait EPA 1999), probably due to high cell biomass. Concentrations of chlorophyll *a* ranged from 1.4 µg l⁻¹ in the non-bloom sites of the mid-Bay to 14.3 µg l⁻¹ in the bloom patches (data not shown). Preceding the bloom, low winds and a calm, stable water-column structure persisted throughout Kuwait Bay.

From examination of the mean concentrations of inorganic nutrients in the months preceding the 1999 Kuwait red tide, it is apparent that just prior to the bloom concentrations of nitrogen increased substantially (Fig. 4A). Ambient nitrogen concentrations during the months of July and August averaged <5 µM, while concentrations in September were nearly double those in previous months. In early October, coincident with the establishment of the bloom, mean concentrations rapidly escalated, such that they were nearly 20-fold those observed in earlier months. A similar comparison for PO₄³⁻ concentrations also revealed an increase from the pre-bloom to early bloom phases; however, the increases were significantly less than those of nitrogen (Fig. 4B). Average concentrations of Si(OH)₄ for the months preceding the bloom varied between 15.8 and 35.6 µM, with higher concentrations occurring in the weeks just before the bloom (data not

shown). During the months of July through August 1999, based on the available monitoring data, the molar ratio of nitrogen:phosphorus was less than that required to balance phytoplankton growth, suggesting limitation by nitrogen. During the early phases of the red tide, with elevated nitrogen, the balance of nutrients was altered to indicate an excess of nitrogen. Consequently, the mean nitrogen:phosphorus ratio of the dissolved nutrients increased nearly 5-fold from August to October (Fig. 4C). Similarly, the mean nitrogen:silicon ratio increased from 0.20 in July to 1.0 in October (Fig. 4D).

During the period October 14 and 15, inorganic nutrient concentrations were not elevated, and concentrations were more similar to those observed in July and August (Fig. 5). This may reflect uptake of these nutrients by the plankton. Data on organic nutrients are not available prior to the sampling of October 14 and 15, but the concentrations during this period suggest a substantial contribution of organic nutrients to the overall nutrient pool (Table 3). On average, DON represented 84% and DOP 64% of the total dissolved nitrogen and phosphorus. Urea concentrations ranged from 0.12 to -1.05 µg at. N l⁻¹, which represented on average ≤2% of the total DON. At the current time, it is not known whether these values are typical for Kuwait Bay.

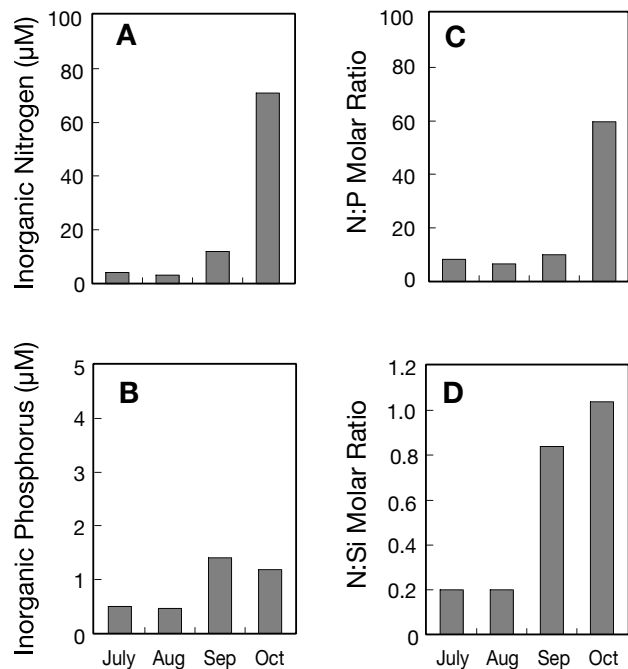


Fig. 4. Mean concentrations of inorganic nitrogen, inorganic phosphorus, molar N:P ratio and molar N:Si ratio in months leading up to the *Gymnodinium* sp. bloom in Kuwait Bay. Data are from long-term monitoring sites from the Kuwait EPA

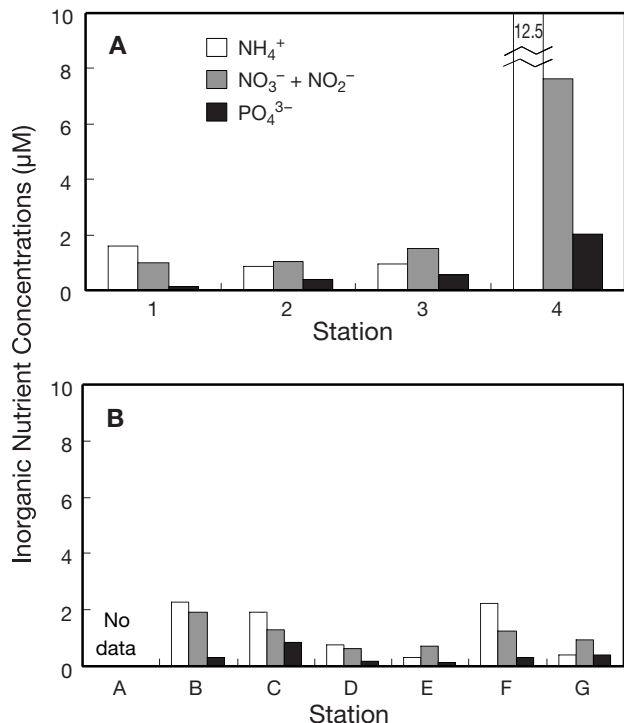


Fig. 5. Inorganic nutrient concentrations for stations sampled on (A) October 14, 1999 and (B) October 15, 1999

DISCUSSION

The October 1999 bloom of *Gymnodinium* sp. in Kuwait Bay is thought to be the first such event leading to fish kills in the Arabian Gulf. Two different organisms were involved in the 'red water' event which was coincident with fish kills in Kuwait Bay—the toxic dinoflagellate *Gymnodinium* sp. and the non-toxic ciliate *Mesodinium rubrum*. The sequence of red water observations and fish kills suggest that the early bloom water discoloration was due to extremely high dinoflagellate concentrations which resulted in fish kills early in the event. Red water observed in the later stages of the bloom after the majority of fish kills occurred (after October 12) could be attributed almost entirely to *M. rubrum* by direct cell counts. Blooms of *M. rubrum* have previously been found to discolor water, possibly

Table 3. Mean (\pm SD) concentrations (μM) of DOC, DON and DOP in samples collected on dates indicated. Also shown is mean percent contribution of urea to the DON pool

Date (1999)	DOC	DOP	DON	%DON as urea
Oct 14	214.8 (11.9)	0.25 (0.13)	20.9 (4.4)	2.0 (1.5)
Oct 15	203.5 (28.4)	0.32 (0.27)	16.2 (5.9)	1.6 (1.0)

due to the release of extracellular products (Sellner 1981, Crawford 1989). Red water was previously observed during a bloom of the non-toxic dinoflagellate *Gyrodinium instriatum* in Kuwait Bay (Subba-Rao et al. 1999); thus the presence of red water in Kuwait Bay is not always indicative of potential fish toxicity.

Red tides and the presence of potentially toxic phytoplankton species in the northwestern Arabian Sea have been previously documented (Subba Rao & Al-Yamani 1998, Subba Rao et al. 1999). Subba Rao et al. (1999) described a bloom of the dinoflagellate *Gyrodinium instriatum* in Kuwait Bay in 1997 that was characterized by high biomass ($>200 \mu\text{g}$ chlorophyll $a \text{ l}^{-1}$) and primary productivity ($>500 \mu\text{g C l}^{-1} \text{ h}^{-1}$), but was not toxic to fish or shellfish. Blooms of the photosynthetic ciliate *Mesodinium rubrum* were also reported for Kuwait Bay in October of 1995, with concentrations up to $160 \mu\text{g}$ chlorophyll $a \text{ l}^{-1}$ and cell concentrations of $1.08 \times 10^6 \text{ cells l}^{-1}$ (Al-Yamani et al. 1997). Other potentially toxic phytoplankton species (e.g. *Pyrodinium bahamense*, *Lingulodinium polyedra*, *Phaeocystis* sp.) have been reported within the northern Arabian Gulf (Dorgham & Mofteh 1989, Al-Yamani et al. 1997, Subba-Rao & Al-Yamani 1998).

Taxonomy of the unarmoured dinoflagellate genus *Gymnodinium* is currently under revision based on electron microscopy and immunological and molecular techniques (Takayama et al. 1998, Hansen et al. 2000, Haywood et al. 2000). This genus is extremely pleomorphic, and significant changes in morphology occur upon preservation due to osmotic sensitivity, often making identification difficult (Steidinger & Tangen 1996, Steidinger et al. 1996, Hansen et al. 2000). Diagnostic characteristics of species are poorly defined and morphological differences between species are often subtle, requiring observation of a large number of cells or electron microscopy. *Gymnodinium* sp., initially described from New Zealand waters (Haywood et al. unpubl.), has also been found in bloom concentrations in the northeastern Gulf of Mexico (K. Steidinger pers. comm.). This species contains the toxin gymnodimine, which is toxic to both mice and fish (Seki et al. 1996) and can produce false positive neurotoxic shellfish poisoning (NSP) results (MacKenzie et al. 1995, 1996). Given the confusion in identification of toxic *Gymnodinium* species (e.g. *Gyrodinium aureolum*, *Gymnodinium mikimotoi*, *G. cf. mikimotoi*) and the description of new, toxic *Gymnodinium* species (Haywood et al. unpubl.), it is likely that blooms of *Gymnodinium* sp. are not restricted to Kuwait Bay, Gulf of Mexico and New Zealand coastal waters and have been misidentified in other areas.

Two alternative hypotheses can explain the observed chronology of the bloom and associated fish kills in Kuwait Bay. Climatic and environmental conditions

may have been such that the bloom developed simultaneously in Iranian and Kuwait waters. Circulation in the northwestern Gulf is an anti-clockwise rotation, however (Hunter 1986, Sheppard et al. 1992, Reynolds 1993). Combined with a report of significant fish kills in Iranian coastal waters almost 2 wk prior to the kills in Kuwait Bay, this suggests that the dinoflagellate may have been present in Iranian waters prior to its appearance in Kuwait Bay, although there are no data currently available to support or refute this. Given current speeds of up to 18.9 cm s^{-1} (Reynolds 1993), transport of bloom populations from Iran to Kuwait coastal waters could have occurred within 2 wk. We also cannot discount the possibility that the Iranian fish kill was completely unrelated to the Kuwaiti event and was due to other causes.

Nutrient measurements made prior to and during the bloom suggest that elevated nutrients, particularly inorganic nitrogen, may have played a role in initiation and/or maintenance of this bloom. Dramatic elevation in inorganic nitrogen and phosphorus levels (up to 20-fold for nitrogen) occurred from late August to early October. The direct source of this nutrient input at this time, however, is unknown. Depletion in silicate relative to nitrogen and phosphorus has also been suggested to favor the outbreak of some harmful bloom species (e.g. Smayda 1989, 1990); however, this was not the case here. In fact, during the bloom, the N:Si ratio was favorable for diatom production.

Of particular note is the high proportion of DON and DOP in Kuwait Bay in October 1999. While data on organic nutrients are not available prior to the October 14 and 15, 1999, sampling, the percent contribution of organic nitrogen and phosphorus to the total dissolved nutrient pool during the bloom exceeded 60%. Recent studies in enriched coastal areas have shown that, while productivity may increase quantitatively with overall nitrogen availability, the DON component may contribute disproportionately to the alteration of phytoplankton succession, leading to the development of harmful algal blooms (Pearl 1988, Berg et al. 1997, LaRoche et al. 1997, Lomas et al. 2001). There are several reasons for this. Many harmful bloom species are mixotrophic or heterotrophic and derive some or all of their energy and nutrition from the uptake of organic molecules or from direct feeding (e.g. Burkholder & Glasgow 1997, Stoecker et al. 1997). It is also possible that nutrient-rich, particularly organic-nutrient-rich, conditions may stimulate other organism(s) that may, in turn, have a stimulatory effect on the harmful bloom species. At this time, there are no available data on the extent to which *Gymnodinium* sp. is or is not mixotrophic.

Two factors previously found to be associated with outbreaks of dinoflagellate blooms are elevated ratios

of DOC:DON and elevated concentrations of urea (Glibert & Terlizzi 1999, Glibert, Magnien, Lomas, Alexander, Fan, Haramoto, Trice & Kana unpubl. data). When *Gymnodinium* sp. dominated the phytoplankton community in Kuwait Bay, the ratio of DOC:DON was approximately twice that of the period when *Mesodinium rubrum* dominated, or when a mixed assemblage was encountered (data not shown). High concentrations of urea have been found to be associated with dinoflagellate blooms in enclosed aquaculture systems (Glibert & Terlizzi 1999). In the Kuwait aquaculture stations, urea was not elevated significantly over other stations, but, overall, the concentrations of urea were relatively high (on average $>0.6 \mu\text{g at. N l}^{-1}$). The potential role of aquaculture in the dynamics of this bloom must be considered, however, as aquaculture ponds and cage culture systems represent an additional source of nutrients to coastal waters. Conditions in and around pens are generally nutrient-enriched, particularly in nitrogen, which can be derived from feed and fertilizer as well as intensified biological transformations (Cho et al. 1996, Burford 1997, Burford & Glibert 1999). The aquaculture industry in Kuwait Bay is relatively recent, with production levels of 500 t yr^{-1} and 600 t of fish in stock (Kuwait Environmental Public Authority 1999). Finfish are a staple of the Kuwaiti diet, and high demands from local markets as well as overexploitation of local fisheries (Al-Mubarak et al. 1999, ROPME 1999) suggest that intensification of aquaculture industry in the region will continue.

Concentrations of both inorganic and organic nutrients measured during the bloom suggest that Kuwait Bay is subject to elevated nutrient loadings from a variety of sources. Industrial and sewage inputs contribute significant inorganic nutrient inputs to the northern Arabian Gulf (ROPME 1999), but their contribution to organic nutrient loading is unknown. The magnitude of nutrient inputs from the relatively recent introduction of aquaculture to the region is also unknown. Iraq has destroyed much of the wetlands area in the region to the northeast and upcurrent of Kuwait at the conjunction of the Tigris, Euphrates and Karoon Rivers and constructed a 'third river' drainage channel which now inputs an additional unknown quantity of pollutants into the northern Arabian Sea (ROPME 1999). Nutrients derived from any one or all of these sources may have provided optimal conditions for bloom initiation and development. The impacts of chronic oil pollution in the area must also be considered, as it has been shown from mesocosm experiments that chronic oil-enrichment enhances phytoplankton biomass and primary production via the reduction of pelagic and benthic grazers (Elmgren et al. 1978). Alterations in trophic dynamics within the region associated with

both chronic and Gulf War related oil pollution may also have contributed to environmental conditions optimal for *Gymnodinium* sp. growth. This toxic algal bloom event highlights the need for monitoring and research programs in the Arabian Sea and Kuwait Bay that focus on nutrients and eutrophication in addition to oil-related pollution issues.

Note added in proof: *Gymnodinium breve*, *G. brevisulcatum* and *G. mikimotoi* were recently transferred to the new genus *Karenia* (Daugbjerg et al. 2000). The *Gymnodinium* sp. that bloomed in Kuwait Bay belongs to this new genus and will be included in it when named (Haywood et al. unpubl.).

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