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1 **Vulnerability and risk of deltaic social-ecological systems exposed to** 2 **multiple hazards**

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4 Michael Hagenlocher^{1,§}, Fabrice G. Renaud^{1,2}, Susanne Haas¹, Zita Sebesvari¹

5

6 ¹ *United Nations University – Institute for Environment and Human Security (UNU-EHS), UN Campus, Platz*
7 *der Vereinten Nationen 1, 53113 Bonn, Germany*

8 ² *University of Glasgow, School of Interdisciplinary Studies, Dumfries Campus, Rutherford/McCowan Building,*
9 *Dumfries DG1 4ZL, UK*

10

11 [§] *Corresponding author; E-Mail: hagenlocher@ehs.unu.edu; Tel: +49-228-815-0250; Fax: +49-228-815-0299*

12

13 **Highlights**

- 14 • Multi-hazard exposure, vulnerability and risk of deltaic social-ecological systems
- 15 • Novel modular, indicator library-based concept and methodology
- 16 • Consideration of hazard-dependent and independent vulnerability indicators
- 17 • Blueprint for vulnerability and risk assessments within and across deltas globally
- 18 • Ecological dimension should be considered more systematically in riskassessments

19

20 **Abstract**

21 Coastal river deltas are hotspots of global change impacts. Sustainable delta futures are increasingly threatened
22 due to rising hazard exposure combined with high vulnerabilities of deltaic social-ecological systems. While the
23 need for integrated multi-hazard approaches has been clearly articulated, studies on vulnerability and risk in
24 deltas either focus on local case studies or single hazards and do not apply a social-ecological systems
25 perspective. As a result, vulnerabilities and risks in areas with strong social and ecological coupling, such as
26 coastal deltas, are not fully understood and the identification of risk reduction and adaptation strategies are
27 often based on incomplete assumptions. To overcome these limitations, we propose an innovative modular
28 indicator library-based approach for the assessment of multi-hazard risk of social-ecological systems across and
29 within coastal deltas globally, and apply it to the Amazon, Ganges-Brahmaputra-Meghna (GBM), and Mekong
30 deltas. Results show that multi-hazard risk is highest in the GBM delta and lowest in the Amazon delta. The
31 analysis reveals major differences between social and environmental vulnerability across the three deltas, notably

32 in the Mekong and the GBM deltas where environmental vulnerability is significantly higher than social
33 vulnerability. Hotspots and drivers of risk vary spatially, thus calling for spatially targeted risk reduction and
34 adaptation strategies within the deltas. Ecosystems have been identified as both an important element at risk as
35 well as an entry point for risk reduction and adaptation strategies.

36

37 **Keywords:** spatial assessment, multi-hazard, index, Amazon delta, Ganges-Brahmaputra-Meghna delta,
38 Mekong delta

39

40 1. Introduction

41 Coastal river deltas are low-lying areas built from sediments transported from upstream and constantly
42 reshaped by the forces of river and tidal water. Within coastal deltas, the patchy transitions between land, fresh,
43 brackish and saline water generate diverse habitats and exceptional biodiversity offering a multitude of
44 ecosystem services (Campbell, 2012). Intensive agricultural production is typically possible thanks to nutrient-
45 rich sediment deposition. Abundant waterways connect settlements and offer opportunities for navigation and
46 transport within the delta and to inland locations. Serving as life support systems, deltas have attracted human
47 settlement for thousands of years and even nurtured the formation of human civilization (Bianchi and Allison,
48 2009; Pennington et al., 2016). Today almost 360 million people live in river deltas (Higgins, 2016) with an
49 average population density of approximately 600 people/km² – an order of magnitude greater than the average
50 population density of the globe (Ericson et al., 2006; Higgins, 2016).

51 Despite considerable advantages, deltaic environments also challenge human activities as these
52 dynamic landscapes (shifting distributaries, erosion and aggradation processes, etc.) tend to have high levels of
53 exposure to multiple (socio-) natural hazards, such as river and tidal flooding, droughts, river bank and coastal
54 erosion, cyclones, storm surges, tsunamis and sea water intrusion (Sebesvari et al., 2016a, 2016b). Some deltas
55 also face additional threats such as earthquakes (Steckler et al., 2008). With rapid urbanization, societal
56 transformation and environmental (including climate) change, it is likely that hazards, vulnerabilities, and

57 associated disaster risk will further intensify in many deltas around the world (Hinkel et al., 2014; IPCC, 2012;
58 Syvitski et al., 2009), undermining sustainable development in these coastal zones of high ecological, economic,
59 and hence societal relevance.

60 Human interventions to (i) tame the dynamic character of the landscape, (ii) exploit natural resources
61 and (iii) reduce exposure to prevailing hazards has led to massive infrastructure development such as canal
62 systems for irrigation and transport, the conversion of wetlands and forests to agricultural, urban and industrial
63 areas (Kuenzer et al., 2014), and the establishment of protective infrastructure to mitigate hazards such as
64 flooding in many deltas around the world. A 52% decline of wetland coverage was recorded between the 1980s
65 and early 2000s in major river deltas (Coleman et al., 2008). Building of upstream dams and reservoirs
66 significantly reduces the amount of sediment reaching the deltas and thus, combined with unsustainable
67 groundwater extraction, contributes to the subsidence of deltas globally (Syvitski et al., 2009). Together with
68 rising sea levels (Horton et al., 2014), significant inundation of land resources and salinization of freshwater
69 resources far inland is projected in many deltas (Giosan et al., 2014). Thus, deltaic landscapes will likely struggle
70 to maintain their structural and functional features (Day et al., 2016; Renaud et al., 2013).

71 Being hotspots of ongoing and projected global change impacts, as well as highly populated,
72 ecologically and economically important sentinels of sustainable development, the state and fate of deltaic
73 social-ecological systems (SES) has great significance locally, nationally and globally. Delta futures will therefore
74 either greatly enhance or detract progress towards achieving the Sustainable Development Goals (SDGs),
75 notably SDG1 (No Poverty), SDG2 (Zero Hunger), SDG6 (CleanWater and Sanitation), SDG11 (Sustainable
76 Cities and Communities), SDG 13 (Climate Action), SDG 14 (Life BelowWater), and SDG 15 (Life on Land)
77 as well as the targets of the Sendai Framework for Disaster Risk Reduction 2015–2030 (Sebesvari et al., 2016a).
78 Assessment of the risks faced by and within deltas based on cross-sectoral and cross-boundary indicators and
79 targets will be key for the identification of targeted risk reduction, resilience-building and adaptation strategies
80 (Chapman and Darby, 2016), and hence the sustainable development of these hotspots for climate change
81 impacts (Szabo et al., 2016a).

82 To date, vulnerability and risk in deltas is mostly characterized with respect to single hazards (Birkmann
83 et al., 2012; Dinh et al., 2012; Mansur et al., 2016), either at the delta scale (Tessler et al., 2015) or based on case
84 studies at the local level (de Andrade et al., 2010; Dinh et al., 2012; Few and Tran, 2010; Islam et al., 2013).
85 Intermediate, sub-delta scale assessments which enable cross-delta comparisons while also delivering planning-
86 relevant information on the scale of sub-delta administrative units are absent (Sebesvari et al., 2016b; Wolters
87 et al., 2016). While the need for considering and analyzing deltas as coupled SES has been increasingly
88 highlighted over the past several years (Brondizio et al., 2016; Day et al., 2016; Szabo et al., 2016b), the majority
89 of existing vulnerability and risk assessments still focus largely on the social and/or economic dimension (Balica
90 et al., 2012; Birkmann et al., 2012; Burton and Cutter, 2008; Chen et al., 2013; Few and Tran, 2010; Mansur et
91 al., 2016; Mondal, 2013; Tessler et al., 2015). Currently integrated assessments focusing on deltas as coupled
92 SES are rare (Lázár et al., 2015; Wolters and Kuenzer, 2015; Sebesvari et al., 2016a, 2016b). As a result,
93 vulnerability and risk in areas with strong social and ecological coupling, such as coastal deltas, is not fully
94 understood and the identification of risk reduction and adaptation strategies are often based on incomplete
95 assumptions.

96 Risk-informed planning of future development as well as targeted disaster risk reduction (DRR) and
97 adaptation strategies and measures (including ecosystem-based options) will not only be important, but
98 existential in deltas in the coming decades. This requires spatially explicit information on the exposure,
99 vulnerabilities and risks associated with different combinations of hazards within deltas in an integrative
100 manner, not only focusing on societal aspects or ecosystems alone, but on interconnected deltaic SES.

101 Addressing the gaps and challenges described above, we present an innovative assessment concept and
102 indicator-based methodology that is sensitive to the specific (multi-) hazard setting in a given delta and can be
103 used as a blueprint for SES-centered spatially explicit vulnerability and risk assessments across and within deltas
104 globally: the Global Delta Risk Index (GDRI). The GDRI presented here uses a modular indicator library-
105 based approach that has been co-developed and piloted with local and regional stakeholders in three globally
106 relevant mega-deltas: the Mekong delta, the Ganges-Brahmaputra-Meghna (GBM) delta, and the Amazon delta.

107 2. Materials and methods

108 The GDRI was constructed building on and extending a multi-step, iterative workflow for index
109 construction and risk assessment (OECD, 2008; Hagenlocher and Castro, 2015; Asare-Kyei et al., 2017). A
110 spatial approach that uses sub-national administrative units was selected to enable cross-delta comparisons of
111 risk while providing information at the sub-delta scale to enable spatially informed local decision-making and
112 intervention planning. The conceptual framework, including major risk components and vulnerability domains,
113 was first defined.

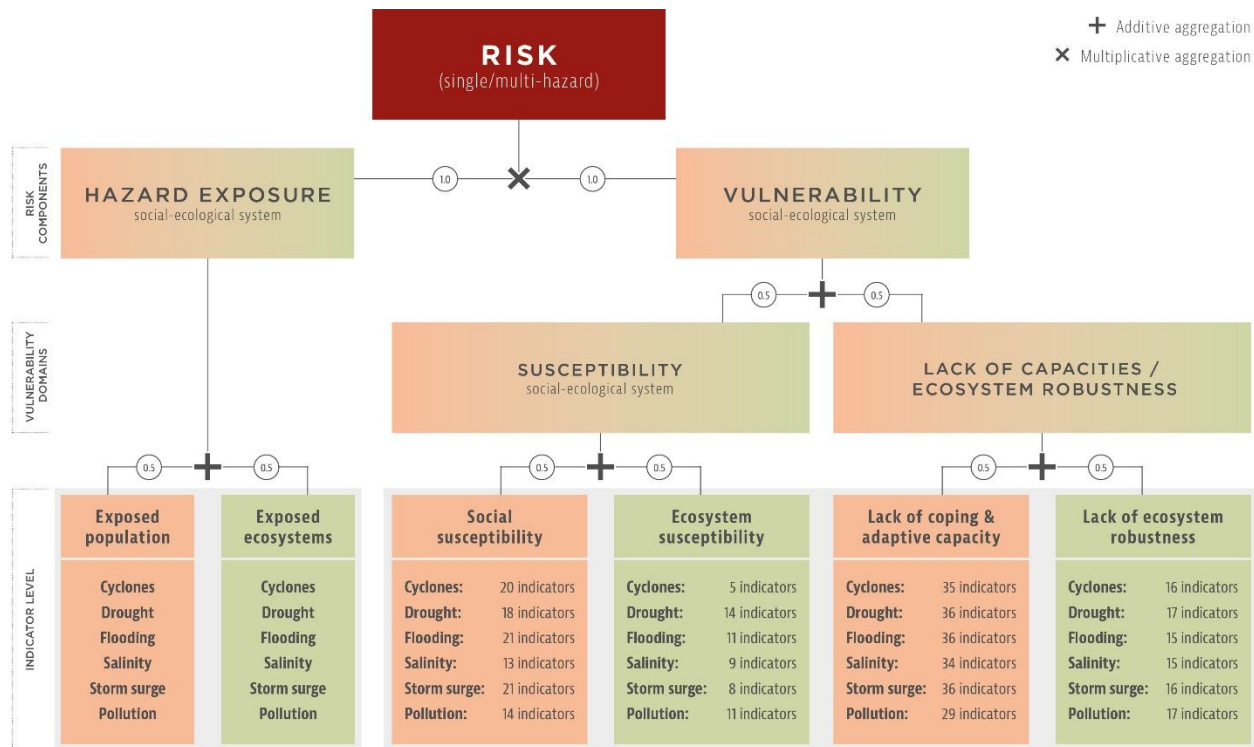
114 2.1. *Framing vulnerability and risk*

115 Over the past decades, conceptual approaches on how to understand risk resulting from human-
116 environment interaction have undergone considerable paradigm shifts from environmentally deterministic,
117 hazard-centric approaches (White, 1973) to political economy and political ecology perspectives (Blaikie et al.,
118 1994; Hewitt, 1983; Lewis, 1999; O'Keefe et al., 1976; Wisner et al., 2004), and finally to holistic concepts that
119 integrate and connect social, economic, political, environmental and governance-related drivers of disaster risk
120 (Birkmann et al., 2013; Eakin and Luers, 2006; IPCC, 2014; Sebesvari et al., 2016b; Turner et al., 2003).
121 Associated theoretical frameworks and models have been instrumental in two main ways: first, in
122 conceptualizing and systematizing vulnerability and risk by sketching out key components and thematic
123 dimensions, and second, by serving as an important first step towards the assessment of vulnerability and risk;
124 for example by guiding the selection of relevant indicators and their combination in a meaningful index.
125 Recognizing the need to consider deltas as coupled SES (Brondizio et al., 2016; Szabo et al., 2016b; Wolters
126 and Kuenzer, 2015) and hence vulnerability of SES (Adger, 2006; Eakin and Luers, 2006; Turner et al., 2003),
127 the GDRI builds on a simplified, inclusive SES-centered vulnerability and risk framework proposed by
128 Sebesvari et al. (2016b) that was adapted from Turner et al. (2003) to fit the idiosyncrasies of deltaic
129 environments. Following conceptualizations by Wisner et al. (2004) and others, in this framework risk is defined
130 as the potential for adverse consequences resulting from the interaction of one or multiple (socio-) natural
131 hazards with vulnerable elements of the SES exposed to these hazards (here operationalized as risk = hazard

132 exposure \times vulnerability). Hazards are operationally defined as the potential occurrence of sudden onset or
133 slow onset natural, anthropogenic or coupled physical events or processes that may cause loss of life, injury, or
134 other health impacts, as well as loss and damage to elements and processes of the SES in place (IPCC, 2014).
135 Exposure refers to the extent to which elements and processes of the SES are in the reach or subjected to one
136 or multiple hazards (IPCC, 2014). Vulnerability is understood as the predisposition of the elements and
137 processes of the SES to be adversely affected (IPCC, 2014) and is composed of four intertwined components
138 in the GDRI: social susceptibility, ecosystem susceptibility, lack of capacity to cope and adapt, and lack of
139 ecosystem robustness (Sebesvari et al., 2016b).

140 Fig. 1 shows the concept and modular structure of the GDRI that was used to assess exposure,
141 vulnerability, and risk of SES to both single or multiple hazards in the three case study deltas, and which –
142 along with an associated ‘library’ of hazard-dependent and independent indicators (Supplementary Material 1)
143 – can be used as a blueprint for further assessments across and within deltaic environments globally. It presents
144 the overall structure and aggregative flow of the GDRI from the indicator level (at the bottom of Fig. 1), to
145 vulnerability domains (i.e. susceptibility of the SES, lack of capacities/ecosystem robustness), risk components
146 (i.e. hazard exposure and vulnerability of the social-ecological system) and final aggregated risk. Following the
147 logic of the modular indicator library, the number of vulnerability indicators varies depending on the hazard.
148 The numbers describe how many indicators are listed for each hazard in the indicator library (Supplementary
149 Material 1).

150



151

152 **Fig. 1.** The Global Delta Risk Index (GDRI): concept and structure. Risk is understood as a function of (i) hazard exposure
 153 and (ii) vulnerability of the SES. It is operationalized applying a modular, flexible indicator set consisting of both hazard-
 154 dependent and independent indicators while being responsive to the specific (multi-) hazard setting in a given delta.
 155

156 **2.2. Indicators and data collection**

157 Relevant hazards and potentially exposed elements of the SES as well as associated hazard-dependent
 158 and independent vulnerability indicators (incl. potential proxy indicators) were identified through a combination
 159 of the outcomes of a systematic review of scientific literature on vulnerability indicators in deltaic SES (detailed
 160 in Sebesvari et al., 2016b) and participatory expert consultations during a series of stakeholder workshops in
 161 the three deltas between 2014 and 2016. The three workshops (one in each delta) were attended by
 162 representatives from local/regional sectoral agencies (i.e. planning, disaster risk management, health,
 163 agriculture, and forestry), academia, and civil society, and aimed at understanding the SES and identifying both
 164 relevant hazards and associated vulnerability indicators in each delta. These indicators were organized in a
 165 modular indicator library (Supplementary Material 1) which distinguishes between hazard-dependent and
 166 independent vulnerability indicators as well as potential proxies. Third, data for the three deltas were acquired
 167 from multiple sources (incl. census data, publicly accessible national and global repositories, as well as directly

168 from scholars who have published on these deltas; see acknowledgements) for the different components of the
169 risk framework: (i) hazard (flooding, cyclones, salinity intrusion, and storm surges), (ii) exposed elements of the
170 SES (i.e. gridded population and land use/land cover data) and (iii) vulnerability of the SES. Since no reliable,
171 comparable data on drought occurrence probability currently exist for the three deltas, probabilistic drought
172 maps were developed for all deltas based on Standardized Precipitation Index (SPI-3) data by using an extreme
173 value modeling approach adapted from Bordi et al. (2007).

174

175 ***2.3. Constructing the Global Delta Risk Index***

176 In constructing the GDRI a spatially explicit approach was pursued. However, due to the paucity of
177 spatially explicit data (i.e. gridded datasets) for many of the indicators, the final calculation of the index was
178 conducted at the sub-national scale using administrative units. The deltas extent dataset published by Tessler et
179 al. (2015) was used to delineate the study area in the three pilot deltas. Exposure of the SES to a single hazard
180 was assessed by calculating the average percentage of both people and ecosystems in hazard-prone areas using
181 gridded data and a spatially explicit approach in a Geographic Information System (GIS). Since people and
182 ecosystems can be exposed to multiple hazards, relative exposure to individual hazards was summed to derive
183 multi-hazard exposure of the SES. To enable cross-delta comparisons of multi-hazard exposure
184 (HAZEXPSES), the final score was divided by five – the maximum number of hazards considered in the
185 analysis. For the vulnerability assessment, following data acquisition, preprocessing of the data was performed
186 using GIS. This includes the calculation of density surfaces (e.g. density of the transportation network,
187 evacuation/shelter places, and emergency services), transformation of absolute into relative values, and
188 calculation of zonal statistics (mean, max) to convert gridded datasets into one score for each administrative
189 unit within the deltas. As a fourth step, an analysis of missing data was performed with no issue observed. Next,
190 potential outliers in the data were examined using both box plots based on the inter-quartile range (i.e. data
191 outside $1.5 \times IQR$) and skewness and kurtosis of the data (Supplementary Material 2). Triangulation was used
192 to verify or falsify potential outliers, and, where relevant, extreme values were treated using a winsorization
193 approach. As a sixth step, multicollinearities within each of the four vulnerability domains were assessed using

194 Kendall's Tau, a non-parametric correlation coefficient that can be used to assess and test correlations between
 195 variables in research contexts with small sample sizes, with $r \geq 0.9$ indicating highly collinear datasets. Statistical
 196 significance was tested using a two-tailed approach. Following this approach, no issue of multicollinearity was
 197 detected (Supplementary Material 2). Both analysis steps were conducted in SPSS (IBM SPSS Statistics). Linear
 198 min-max normalization was applied to rescale the indicators to a range between zero and one. To render the
 199 datasets comparable, the lowest min and highest max values across the three deltas were used in the
 200 normalization process for all three deltas. Indicators where high scores would contribute to reduced
 201 vulnerability and risk (see indicator library; Supplementary Material 1) were inverted during the normalization
 202 process. In step eight, the pre-processed and normalized indicators (x_i) were combined into the four
 203 vulnerability domains (VD ; i.e. ecosystem susceptibility, social susceptibility, lack of ecosystem robustness, and
 204 lack of coping/adaptive capacities) using additive aggregation (Eq. (1)). Following the modular structure of the
 205 indicator library, only those indicators were considered that are relevant for the specific (multi-)hazard setting
 206 in the three deltas as defined by stakeholder input.

207

$$208 \quad VD = \sum_{i=1}^n (w_i * x_i') \quad (1)$$

209

210 While the design of the GDRI (Fig. 1) principally enables users to specify weights (w_i) for each
 211 normalized indicator (x_i'), for example through expert consultation, equal weights were applied for the
 212 construction of the GDRI in the three case study deltas. To analyze vulnerability domains of the SES (VD_{SES}),
 213 social and ecosystem susceptibility were aggregated into a metric representing susceptibility of the SES while
 214 lack of coping/adaptive capacities was combined with lack of ecosystem robustness into a metric representing
 215 the lack of capacities/robustness of the SES and applying equal weights (Eq. (2)).

216

$$217 \quad VD_{SES} = \sum_{j=1}^n (w_j * VD_j) \quad (2)$$

218 The two vulnerability domains of the SES, i.e. susceptibility of the SES (VD_{SES1}) and lack of capacities
219 and robustness of the SES (VD_{SES2}), were then further combined into an index score representing vulnerability
220 of the SES (VU_{SES}) using the average of the two (Eq. (3)).

221

$$222 \quad VU_{SES} = \left(\frac{VD_{SES1} + VD_{SES2}}{2} \right) \quad (3)$$

223

224 Ultimately, following the modular framework, (multi-) hazard exposure and vulnerability of the SES
225 were combined in a (multi-hazard) risk index ($RISK_{SES}$) through multiplicative aggregation, whereby both risk
226 components are weighted equally. Thereby, the hazard component of the framework was indirectly considered
227 in the exposure term of the equation following existing risk assessment approaches (Hagenlocher and Castro,
228 2015; BEH and UNU-EHS, 2016; Wannewitz et al., 2016). Aggregation of indicators and domains was carried
229 out in Excel and results visualized in a GIS. Results are presented in Fig. 2.

230

$$231 \quad RISK_{SES} = HAZEXP_{SES} \times VU_{SES} \quad (4)$$

232

233 **2.4. Reliability analysis**

234 Drawing on a metric developed in the context of the Index for Risk Management (INFORM;
235 <http://www.inform-index.org>), a reliability index was developed and applied to each of the three deltas to
236 increase transparency regarding the quality of data used to calculate the GDRI. Adjusted from the INFORM
237 reliability metric, it takes into account the following criteria: (i) percentage of imputed data across indicators,
238 (ii) percentage of hazard data available, (iii) percentage of proxy indicators, (iv) percentage of indicators at
239 provincial level (here assumed to be positive, although higher spatial resolution would be desirable), and (v)
240 recency (average age) of the data. The corresponding scores for each criteria were rescaled to values between
241 zero and one, aggregated, and averaged in order to derive the final reliability scores. In this context a score of
242 1 represents high reliability and a score of 0 implies low reliability. Since the GBM delta is located both in India

243 and Bangladesh, the reliability index was computed for each country separately (Supplementary Material 2). All
244 analyses were performed using Microsoft Excel.

245 **2.5. Data availability**

246 The data that support the findings of this study are provided as Supplementary Data 1. Supplementary
247 Material 3 provides an overview of the datasets and sources for each of the indicators that were used to
248 construct the GDRI.

249 **3. Results**

250 **3.1. Multi-hazard risk of deltaic social-ecological systems**

251 Using the GDRI framework and approach, we assess single and multi-hazard exposure, vulnerability,
252 and risk of SES across and within the three deltas (Table 1). While deltas are generally confronted with a number
253 of climate and non-climate related hazards, the assessment focuses only on those hazards that were mentioned
254 as relevant by the local stakeholders during the workshops in the respective deltas, i.e. flooding, drought, and
255 salinity (for all deltas) as well as cyclones and storm surges (only for the GBM delta) and where data were
256 available. We show that average multi-hazard risk is highest in the GBM delta (0.21 in a range from 0 to 1) and
257 lowest in the Amazon delta (0.09 in a range from 0 to 1) with the largest within-delta variability in the GBM
258 delta. The latter effect is partially strengthened by the fact that the GBM is shared among two distinct
259 neighboring countries – India and Bangladesh. On average, multi-hazard risk is higher in the Bangladeshi
260 portion (risk score: 0.22) as compared to the Indian portion of the GBM delta (risk score: 0.14). Being prone
261 to multiple hazards (i.e. flooding, droughts, salinity intrusion, cyclones and storm surges), SES exposure is
262 highest in the GBM delta, followed by the Mekong delta which features high exposure to flooding in the
263 upstream parts of the delta, high exposure to droughts in the southern parts of the delta and high exposure to
264 salinity intrusion along the coastline. Multi-hazard exposure is lowest in the Amazon delta, where primarily the
265 southeastern municipalities are affected by both flooding and droughts (Table 1).

266

267 **Table 1**

268 Multi-hazard risk, hazard exposure, and vulnerability of the SES across the three deltas.

	Multi-hazard risk of SES		Multi-hazard exposure of SES		Multi-hazard vulnerability of SES		Multi-hazard susceptibility of SES		Multi-hazard lack of capacities / robustness of SES	
	Mean	<i>SD</i>	Mean	<i>SD</i>	Mean	<i>SD</i>	Mean	<i>SD</i>	Mean	<i>SD</i>
Amazon	0.09	<i>0.04</i>	0.17	<i>0.08</i>	0.51	<i>0.02</i>	0.42	<i>0.04</i>	0.59	<i>0.03</i>
GBM	0.21	<i>0.12</i>	0.34	<i>0.19</i>	0.62	<i>0.04</i>	0.50	<i>0.04</i>	0.73	<i>0.05</i>
Mekong	0.16	<i>0.05</i>	0.31	<i>0.11</i>	0.51	<i>0.02</i>	0.42	<i>0.02</i>	0.61	<i>0.02</i>

269

270 Disaggregating the highly aggregated index, the analysis further reveals major differences between
 271 social and environmental vulnerability across the three deltas, notably in the Mekong and the GBM deltas where
 272 environmental vulnerability is significantly higher than social vulnerability. Environmental vulnerability is
 273 highest in the GBM and lowest in the Amazon delta while social vulnerability is highest in the GBM and lowest
 274 in the Mekong delta. In all deltas, social and ecological susceptibility scores are lower than those representing
 275 the lack of short-term coping and long-term adaptive capacities or the lack of ecosystem robustness (Table 2).

276

277 **Table 2**

278 Decomposing multi-hazard vulnerability of SES across the three deltas.

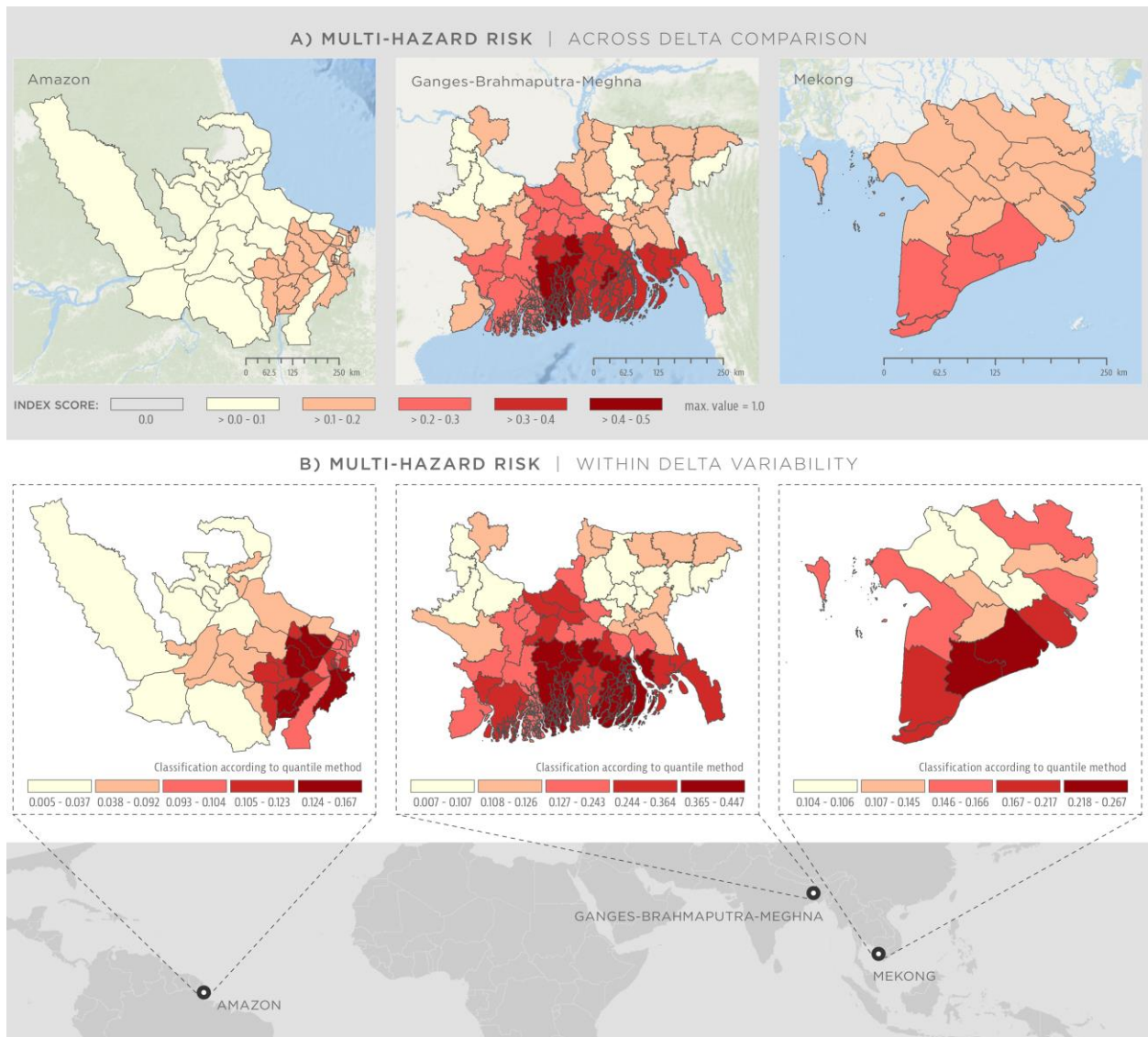
	Social vulnerability (multi-hazard)	Environmental vulnerability (multi-hazard)	Social susceptibility (multi-hazard)	Ecosystem susceptibility (multi-hazard)	Lack of coping & adaptive capacities (multi-hazard)	Lack of ecosystem robustness (multi-hazard)
	Mean	Mean	Mean	Mean	Mean	Mean
Amazon	0.52	0.49	0.47	0.37	0.57	0.61
GBM	0.56	0.67	0.40	0.60	0.71	0.74
Mekong	0.37	0.66	0.29	0.54	0.44	0.78

279

280 Key drivers of ecosystem susceptibility in all deltas include freshwater scarcity and low soil organic
 281 matter. Additionally low forest connectivity and low species richness are important drivers of ecosystem
 282 susceptibility in the GBM and Mekong deltas, whereas additional relevant drivers include the likelihood for
 283 groundwater pollution (probability of arsenic in groundwater) in the GBM, and low cation exchange capacity
 284 of the soil in the Amazon delta respectively. Low coverage by protected areas is an important factor influencing

285 the lack of ecosystem robustness in each of the deltas. Additionally, decreased ecosystem functionality and
286 species abundance are important drivers in the GBM and Mekong deltas.

287 Fig. 2 shows spatial multi-hazard risk patterns and variability for SES both across (Fig. 2, panel A) and
288 within (Fig. 2, panel B) the three deltas at the sub-delta scale (i.e. for 49 municipalities in the Amazon delta, for
289 59 districts in the GBM delta and for 13 provinces in the Mekong delta). Yellow represents lower multi-hazard
290 risk, while darker shades of red indicate higher levels of multi-hazard risk.

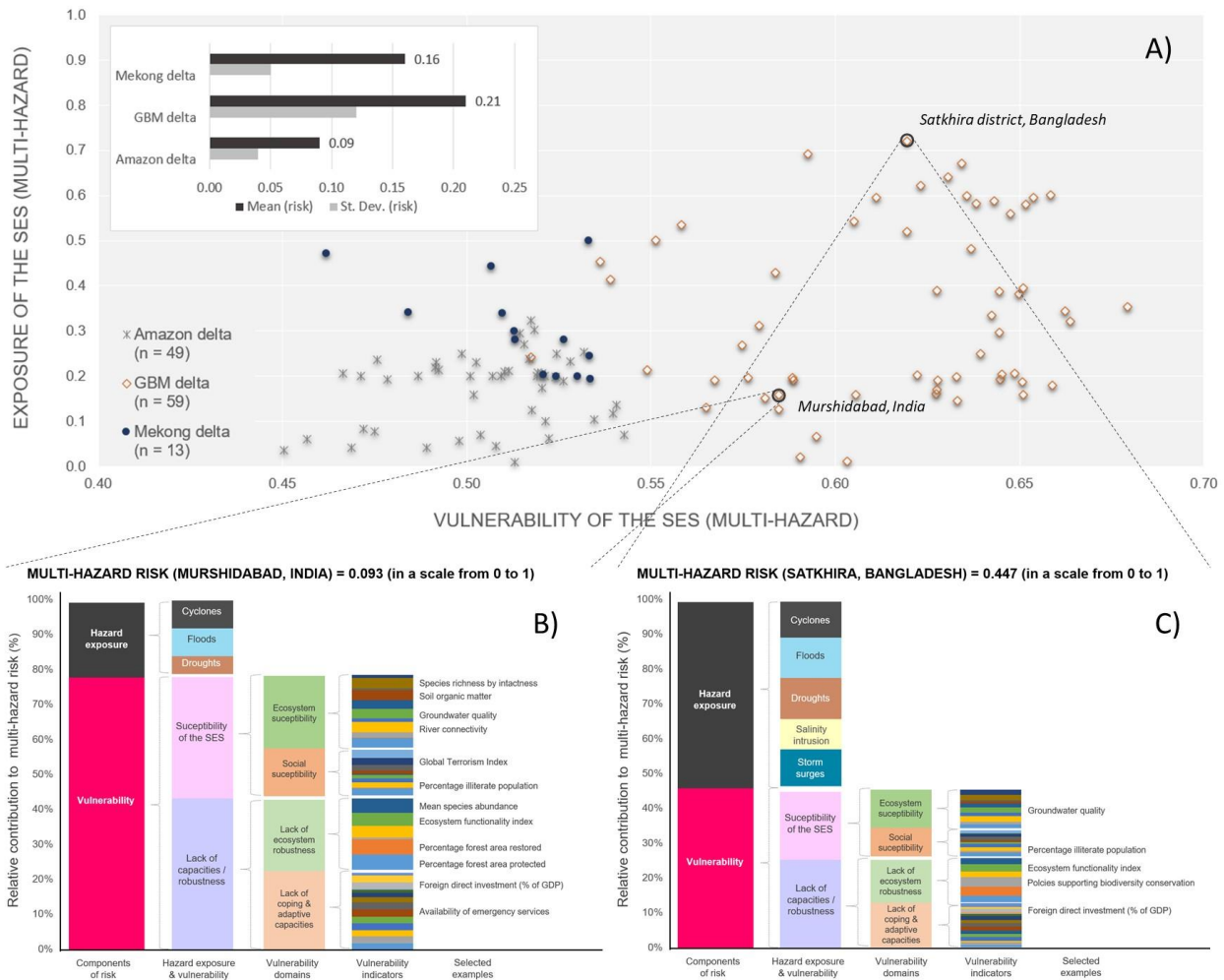


291

292 **Fig. 2.** Multi-hazard risk of the social-ecological system: (A) across and (B) within deltas. Note: Visualization in Panel B
293 is based on quantiles, i.e. colors are not comparable across deltas.

294

295 The variability of the main components of multi-hazard disaster risk – hazard exposure and
 296 vulnerability of the SES – is visualized in Fig. 3 for each of the sub-delta administrative units, with the greatest
 297 variability in the GBM delta. This highlights that differences in multi-hazard risk across and within the three
 298 deltas are largely driven by variations in exposure given that overall vulnerability of the SES is similarly high
 299 (between 0.51 and 0.62) in all deltas. However, the risk profiles show that the drivers of vulnerability, despite
 300 similar overall vulnerability scores, vary between sub-delta regions.



301

302 **Fig. 3.** Risk profiles in the Amazon, GBM, and Mekong delta: based on variability in multi-hazard exposure and
 303 vulnerability of SES across deltas (Panel A). Panels B and C show the relative contribution of the risk components,
 304 vulnerability domains and indicators to the final risk scores for two selected districts in the GBM delta, constituting
 305 illustrative risk profiles.

306

307 **3.2. Reliability**

308 On a scale from 0 (lowest reliability) to 1 (highest reliability), reliability scores of 0.74, 0.79 and 0.81
309 were calculated for the Amazon, the GBM and the Mekong Delta, respectively. The assessment shows that the
310 risk estimates of all deltas were calculated based on relatively recent data with an average age of around 3 years
311 but suffered from a lack of data available at the provincial level and relatively frequent use of proxy indicators
312 (Supplementary Material 2).

313 **3.3. Indicator library**

314 Relevant indicators representing vulnerability of deltaic SES were added to an indicator “library”
315 (Supplementary Material 1) that can support future assessments of vulnerability and risk in deltas. In the library
316 indicators are organized according to the four main vulnerability domains (social susceptibility, ecosystem
317 susceptibility, lack of coping and adaptive capacities, lack of ecosystem robustness) in a modular structure, i.e.
318 according to their relevance for the six different (socio-)natural hazards relevant in the three deltas considered
319 in this analysis: (i) cyclones, (ii) drought, (iii) floods (pluvial/fluvial), (iv) salinity intrusion, (v) storm surges, and
320 (vi) pollution. The library can easily be extended if applied in different delta settings, for example by considering
321 additional hazards. The library comprises indicators that can be represented in a spatially explicit manner (e.g.
322 deforestation, wetland loss, forest connectivity, or travel time to the closest city), at the level of administrative
323 units (e.g. illiteracy, dependency ratio, or poverty), as well as indicators on policy processes that affect an entire
324 delta (e.g. existence of disaster risk management and/or adaptation strategies, integrated development plans, or
325 policies contributing to biodiversity conservation or coastal protection). Distinguishing between hazard-
326 dependent and independent indicators, the modular approach is responsive to the specific (multi-)hazard setting
327 of a given delta, enabling developers of future (multi-hazard) vulnerability and risk assessments to include only
328 those indicators that are relevant for a particular hazard configuration within the delta. Next to the modular set
329 of indicators, a list of potential proxy indicators – that can be used when appropriate data to represent the
330 actual vulnerability indicators do not exist (e.g. when conducting assessments in data scarce environments) – is
331 proposed. These proxy indicators were identified during the review of scientific literature by Sebesvari et al.

332 (2016b), during the stakeholder workshops in the three deltas, and by analyzing correlations (in SPSS software)
 333 between indicators where data were available for the three deltas. Table 3 provides an overview of the number
 334 of indicators in each vulnerability domain in the indicator library as well as the final number of indicators that
 335 were included in the assessment considering data availability in the three deltas.

336

337 **Table 3**

338 Number of vulnerability indicators relevant for specific hazards (library vs. used in the assessment).

	Social susceptibility		Ecosystem susceptibility		Lack of coping capacity		Lack of adaptive capacity		Lack of ecosystem robustness	
	Library	Used *	Library	Used *	Library	Used *	Library	Used *	Library	Used *
Flooding ^a	21	13 (62%)	11	8 (73%)	25	16 (64%)	11	3 (27%)	15	8 (53%)
Droughts ^a	18	11 (61%)	14	10 (71%)	23	15 (65%)	13	3 (23%)	17	8 (47%)
Storms ^b	20	13 (65%)	5	3 (60%)	24	16 (67%)	11	3 (27%)	16	8 (50%)
Storm surges ^b	21	13 (62%)	8	4 (50%)	24	16 (67%)	11	3 (27%)	16	8 (50%)
Salinity intrusion ^c	13	9 (69%)	9	5 (56%)	21	13 (62%)	13	3 (23%)	15	7 (47%)
Pollution ^d	14	0 (0%)	11	0 (0%)	19	0 (0%)	10	0 (0%)	17	0 (0%)
Total number	23	13 (57%)	16	11 (69%)	26	16 (62%)	13	3 (23%)	17	8 (47%)

339 * to enable comparability across the three deltas, indicators were only included in the assessment if data was available
 340 for all three deltas (the Amazon delta, the GBM delta, and the Mekong delta). During the data acquisition process it
 341 became evident that data availability also varies across deltas. ^a relevant in all three deltas; ^b relevant only in the GBM
 342 delta; ^c relevant in all three deltas, but data only available for the GBM and Mekong delta; ^d relevant in all three deltas,
 343 but no data available for all deltas.

344

345 4. Discussion and further research

346 Despite the progress that has been achieved in DRR at national and local levels over the past several
 347 decades, disasters resulting from the interaction of (socio-) natural hazards with increasing exposure and high
 348 levels of vulnerability of SES continue to undermine sustainable development and pose a challenge to achieving
 349 the SDGs. As a consequence, the Hyogo Framework for Action (2005–2015) and the more recently adopted
 350 Sendai Framework for Disaster Risk Reduction (2015–2030; UNISDR, 2015) call for a better understanding of

351 disaster risk (priority 1 in the Sendai agreement) and enhanced efforts to reduce exposure and vulnerability and
352 build resilience at all levels using inclusive, multi-hazard approaches. Next to identifying drivers of disaster risk,
353 this also requires the assessment and monitoring of disaster risk. Despite the clearly articulated urgency for
354 assessments in the global DRR agenda and the recognition of the need for assessing the vulnerability of SES
355 (Turner et al., 2003; Eakin and Luers, 2006; Fuchs et al., 2012), notably in coastal deltas that are shaped by
356 intensive human-environment interaction and exposure to multiple hazards (Brondizio et al., 2016; Sebesvari
357 et al., 2016b), to our knowledge no approach exists to date that can be used for the spatial assessment and
358 monitoring of vulnerability and risk of SES in a multi-hazard setting. The presented GDRI addresses these
359 gaps, and constitutes a novel, transparent, indicator library-based concept and methodology for the integrated
360 assessment of single or multi-hazard vulnerability and risk across and within deltaic SES. Our results confirm
361 the need emphasized by De Lange et al. (2010) to systematically consider indicators of environmental
362 vulnerability when assessing disaster risk and identifying risk reduction and adaptation strategies, while
363 reemphasizing the need for enhanced efforts to sustainably manage, conserve, and restore ecosystems and their
364 services. Through its modular design, the GDRI also overcomes the limitations of existing multi-hazard risk
365 assessment approaches that do not differentiate between hazard-dependent and hazard-independent indicators
366 (Greiving, 2006; Liu et al., 2013; BEH and UNU-EHS, 2016; Wannewitz et al., 2016) – a need that has been
367 underscored in recent reviews of multi-hazard risk methodologies for natural hazards (Kappes et al., 2012;
368 Gallina et al., 2016). Going beyond case studies at the local level (de Andrade et al., 2010; Birkmann et al., 2012;
369 Dinh et al., 2012; Islam et al., 2013) and global assessments that do not capture differences in vulnerability and
370 risk within deltas (Tessler et al., 2015), the sub-delta scale applied here enables the identification of hotspots
371 and variability of risks within deltas. It is thus in line with the need to “develop, periodically update and
372 disseminate, as appropriate, location-based disaster risk information, including risk maps, to decision makers,
373 the general public and communities at risk” as articulated in the Sendai agreement (UNISDR, 2015, p.15).

374 As pointed out by de Sherbinin et al. (2017, p.415), “indicators and indices can help to reduce
375 complexity in policy-relevant ways, providing an important link between science and policy and helping to point
376 decision-makers towards potential solutions to problems at the human–environment interface.” This has also

377 been highlighted by Abson et al. (2012, p. 523) who conducted a study on vulnerability of SES in the Southern
378 African (SADC) region and concluded that from a policy perspective “contextualized, ‘information-rich’
379 vulnerability indices can prove useful as they provide a compromise between the rich and difficult to interpret
380 detailed information provided by a large suite of individual vulnerability indicators and easy to visualize, but
381 potentially ‘information poor’, aggregate vulnerability indices.” This is particularly relevant in the adaptation
382 arena, where millions of dollars are already being spent or will be spent in the decade to come, and where spatial
383 prioritization is key (de Sherbinin et al., 2017). The presented approach can inform delta development planning,
384 such as those strategic plans published for the Mekong delta (Government of Vietnam and Dutch Government,
385 2013) and the GBM delta (<http://www.bangladeshdeltaplan2100.org>) by highlighting the fact that different
386 regions of a delta are exposed to different hazards or combinations of hazards and have consequently different
387 vulnerabilities and risks. Decomposing the GDRI into its risk components (hazard exposure and vulnerability),
388 vulnerability domains (susceptibility of SES and lack of capacities/robustness of SES), and their underlying
389 indicators (Fig. 3) for use by decision makers, practitioners, and analysts, supports the identification and
390 definition of spatially targeted risk reduction and adaptation strategies and/or measures. Web-based approaches
391 that enable an interactive visualization and querying of the results and associated risk profiles can be
392 instrumental for such purposes (Kienberger et al., 2013) and will be compatible with the modular structure of
393 the GDRI. The associated indicator library enables the further transferability of the presented concept and
394 methodology to deltas globally. Depending on the specific hazard configuration in the relevant delta(s), the
395 library provides guidance on which vulnerability indicators or proxies to take into account, while their site-
396 specific weight can be evaluated through expert consultations. However, the library is open to extensions in the
397 future and could also be adapted to different spatial scales and/or hazard configurations, such as earthquakes,
398 tsunamis, etc. Underscoring the need to consider coastal river deltas as closely coupled human-environmental
399 systems and providing a blueprint for future assessments of the vulnerability and risks faced by these systems
400 that has been shaped in close collaboration with local stakeholders in the three case study deltas, the paper also
401 responds to some of the fundamental research questions in sustainability science as reported by Kates (2011),
402 e.g. on what determines the vulnerability of human-environment systems.

403 Despite these advantages, the GDRI and its application to the three case study deltas has several
404 limitations which need to be addressed in future research. First, the index-based assessment is static in time,
405 space as well as in regards to potentially non-linear linkages between SES indicators. For example it does not
406 enable evaluating the effect of interventions in one administrative unit of a delta (e.g. in terms of infrastructure
407 to control water, etc.) on risk levels in another location of the delta. Second, data for the indicators were drawn
408 from different sources and for different years, with implications on the reliability of the results. The presented
409 reliability index makes this weakness transparent. Third, data quality, data availability, the paucity of spatial (i.e.
410 gridded) datasets, and the use of proxy indicators can be considered a limiting factor – particularly in data-
411 scarce environments. By highlighting this limitation (Table 3), this paper also makes a call to governments for
412 enhanced collection, management and provision of updated, reliable spatial data. This is fundamental if Priority
413 1 of the Sendai agreement (“Understanding disaster risk”) is to be met with success. Although the GDRI is
414 designed to pursue a spatial approach, the lack of spatially explicit, gridded data for many of the indicators that
415 were considered in the analysis presents another limitation. While the exposure analysis enables planners to
416 understand where within a delta people or ecosystems are located relative to hazard zones, the GDRI was
417 constructed at administrative sub-delta scale and does not enable planners to make any inferences on
418 vulnerability or risk of SES within these units. Fourth, an associated limitation is that while using sub-national
419 administrative units for the communication of the results supports decentralized decision-making, the coarse
420 resolution applied in this study (Amazon delta: municipal level; GBM delta: district level; Mekong delta:
421 province level) does not capture potential urban-rural differences in disaster risk within these units. Fifth, the
422 linear min-max normalization approach applied here, although most commonly used in index construction and
423 vulnerability assessment (Beccari, 2016), presumes that for each unit increase between zero and one there is a
424 proportional and corresponding increase in vulnerability which is often not the case. Using expert-defined
425 thresholds and/or value functions for each indicator could help to at least partly overcome this issue – albeit
426 introducing another layer of subjectivity and complexity in risk assessments. Sixth, the relevance of vulnerability
427 indicators could vary spatially across and within deltas. While the GDRI has been principally designed to enable
428 users to incorporate indicator weights (see Eq. (1)), equal weights were used for all indicators due to the lack

429 of evidence on their location-specific relevance. Hagenlocher and Castro (2015) compared the impact of
430 different weighting schemes, including weights based on regression coefficients, principal component analysis
431 and equal weights in the context of a spatially explicit grid-based vulnerability and risk assessment and found
432 no statistically significant impact of the weighting schemes. Nonetheless, going beyond the reliability metric
433 presented here, follow-up research should analyze the sensitivity of the index to weighting choices. Lastly, the
434 hierarchical design of the GDRI multiple aggregation steps for assessing multi-hazard SES vulnerability poses
435 another challenge: the more times indicator scores are incrementally averaged, the less variability there is in the
436 output scores.

437 To face the above challenges, attempts are currently underway to dynamize the GDRI using a Bayesian
438 modeling approach and to further downscale the GDRI to the sub-municipal level for selected municipalities
439 in the Amazon delta. This will also provide an opportunity to validate the findings against existing loss and
440 damage information – a topic that has insufficiently been addressed to date. To test the transferability of the
441 GDRI and its associated indicator library, as well as its concordance with other index-based approaches, an
442 assessment of multi-hazard exposure, vulnerability and risk at the census tract level is currently ongoing for the
443 Mississippi delta in the United States.

444 **5. Conclusions**

445 The aim of this paper is to present the development of a novel approach that can be used as a blueprint
446 for spatial single- or multi-hazard risk assessments of closely coupled deltaic social-ecological systems at sub-
447 delta scales and apply it to the Amazon, Ganges-Brahmaputra-Meghna, and Mekong deltas. Novelty includes
448 the introduction of a modular risk assessment approach that differentiates between hazard-dependent and
449 independent vulnerability indicators, the introduction and piloting of an indicator library that can provide
450 guidance for future assessments in river deltas, and the application of a social-ecological systems perspective in
451 the assessment.

452 Findings from this study underscore the importance of not only considering societal vulnerabilities as
453 a constituent of risk, but also including the ecological dimension more systematically when conducting risk

454 assessments to inform disaster risk reduction and climate change adaptation planning and strategies – notably
455 in places of strong social-ecological coupling such as coastal river deltas.

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466 **Appendix A. Supplementary data**

467 Supplementary data to this article can be found online at <https://doi.org/10.1016/j.scitotenv.2018.03.013>.

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